1	Environmental and crown related factors affecting street tree transpiration
2	in Helsinki, Finland
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12 Abstract

13	We investigated the drivers of street tree transpiration in boreal conditions, in order to better
14	understand tree water use in the context of urban tree planning and stormwater management. Two
15	streets built in Helsinki in 2002, hemiboreal zone that had been planted either with $Tilia \times vulgaris$
16	or Alnus glutinosa f. pyramidalis were used as the study sites. Tree water use was measured from
17	sap flow over the 2008-2011 period by the heat dissipation method. Penman-Monteith based
18	evapotranspiration models of increasing complexity were tested against the tree water use
19	measurements to assess the role of environmental and tree related factors in tree transpiration.
20	
21	Alnus and Tilia respectively used 1.1 and 0.8 liters of water per m^2 of leaf area per day under ample
22	water conditions, but the annual variation was high. The Penman-Monteith evapotranspiration
23	estimate and soil water status changes explained over 80% of the variation in tree transpiration
24	when the model was parameterized annually. The addition of tree crown surface area in the model
25	improved its accuracy and diminished variation between years and sites. Using single
26	parameterization over all four years instead of annually varying one did not produce reliable
27	estimates of tree transpiration. Tree transpiration, scaled to different canopy cover percentages,
28	implied that the columnar Alnus trees could transpire as much as all annual rainfall at or less than
29	50% canopy cover.
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31 Keywords: evapotranspiration, street trees, urban, water use

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33 Introduction

Knowledge of the water use of street trees is useful, for example, as a basis for soil resource
allocation when planning tree-lined streets. Increasingly urban trees are also considered as a
component of integrated urban stormwater management, thus an understanding of their actual and

potential water use and the environmental and tree related factors affecting tree water use areneeded.

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40 Measuring the water use of urban trees is challenging due to their open spatial arrangements with 41 specific local environments in comparison to natural canopies. Direct transpiration measurements 42 by using micrometeorological methods available for homogeneous canopies are not necessarily well 43 suited for solitary or openly spaced urban trees with irregular spatial arrangements (Lee 2000, 44 Villegas et al. 2014). On the other hand, accurate up-scaling of leaf level transpiration 45 measurements to crown and canopy level can be difficult to achieve (Jarvis 1995) due to e.g. the 46 heterogeneous but non-random distribution of trees and also to the urban constructions in close 47 proximity to the crowns of interest. Between these scales of local and leaf level measurements, the 48 measurements of sap flow (Granier 1985) as a mean of estimating whole-tree transpiration rates are 49 readily usable for even solitary trees. However, also this method can be problematical due to e.g. 50 nonuniform sap flow within the stem (Clearwater et al. 1999), particularly with large trees. 51 Nonetheless, sap flow measurements are among the best, if not the only, practical method to 52 estimate the effects of climate and soil conditions on urban tree transpiration (e.g. Litvak et al. 53 2011, McCarthy and Pataki 2010).

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Water use measurements of urban trees made *in-situ* have been published only for relatively few species and climate regions and commonly over short time periods (Chen et al. 2011, Litvak et al. 2011, McCarthy and Pataki 2010, Pataki et al. 2011, Peters et al. 2010, Wang et al. 2011). The relevance of water use measurements for natural and agricultural ecosystems in comparison to the estimation of urban tree water use is complicated by different environments and species composition. The interactions of tree function and environment in urban conditions differ from those observed in the natural habitats of trees. For example, due to commonly observed higher air

temperatures and lower relative humidity, water vapor pressure deficit can be substantially higher in
urban areas, and soil water availability may be limited when compared to natural surroundings
(Nielsen et al. 2007). Such conditions can lead to stomatal regulation characteristics that are
different from native environments (McCarthy and Pataki 2010, Litvak et al. 2011).

66

67 The Penman-Monteith (PM) equation is commonly used to estimate potential evapotranspiration 68 (ET₀) from vegetated surfaces (Monteith 1965; Penman 1948). ET₀ considers the influence of 69 meteorological variables on transpiration. The PM equation or model is widely used in an 70 agricultural context and to estimate irrigation needs for nursery trees (e.g. Beeson 2012, Grant et al. 71 2009). In these estimations, the obtained ET_0 value is commonly multiplied by an empirically 72 derived crop coefficient k_c (Doorenbos and Pruitt 1977) or other scaling factor that summarizes the 73 influence of vegetation properties. This agriculture-based approach to determine tree water use does 74 cause some problems when applied to street trees. For example, the canopy in the urban 75 environment is often not closed, which renders the use of a general crop coefficient problematic 76 (Beeson 2012, Hagishima et al. 2007), and also soil water content may limit transpiration (Bernier 77 et al. 2006). Describing surface resistance or canopy conductance properties are more important at 78 both canopy and leaf level and need to be integrated into the model in more detail (Granier et al. 79 2000). On the other hand, the PM model is fairly robust and the application of the PM derived ET_0 80 is widespread in irrigation management, and is a familiar approach for many who work in urban 81 greening industry. Models that are used to describe and predict the transpiration of forest trees (e.g. 82 Granier et al. 2000, Medlyn et al. 2007, Mäkelä et al. 2006) are an obvious alternative to the PM 83 model. These models are also often based on the PM approach, but are more detailed, for example, 84 in describing the transpiration control by the trees, and require input variables such as 85 photosynthesis and respiration parameters or radiation components that may not be readily available 86 in an urban setting. Also in these models a continuous canopy is often assumed.

88	In this study, we measured and modelled street tree water use in boreal conditions in Helsinki,
89	Finland. The aim was to assess the responses of urban tree transpiration to environmental variables
90	and to examine the magnitude of tree water use in relation to annual precipitation. More
91	specifically,
92	1. Street tree sap flow was measured to attain estimates of tree water use over a four-year period.
93	2. A suitable version of PM model was developed to best describe the street tree sap flow. The
94	details of canopy properties description were varied and effects of these changes on the model
95	performace assessed.
96	3. The most important environmental variables driving tree sap flow were sought based on PM-
97	model fit against sap flow observations. Additionally, an analysis of model residuals was performed
98	against environmental variables, and sap flow observations were regressed directly aganist the
99	environmental variables.
100	4. PM models were applied to estimate annual tree water use and used to compare tree transpiration
101	against precipitation in an urban context assuming different canopy covers.
102	
103	Materials and methods
104	Study sites and measurements
105	The study was conducted in two experimental streets situated close to the University of Helsinki
106	Viikki campus, about 9 km from the Helsinki city centre (N60°15', E25°03') (Figure 1). The two
107	streets were planted with 15 Tilia × vulgaris Hayne trees and 22 Alnus glutinosa (L.) Gaertn. f.
108	pyramidalis 'Sakari' trees, respectively, in late 2002. From now on the two streets are referred as
109	Tilia and Alnus site. The growing media consisted of three pre-mixed structural soils in both sites,
110	and these ranged from 32-64 mm to 64-150 mm in stone size and 65-70% in stone matrix volume.
111	The soil plot dimensions for each tree were 3 m wide, 1 m deep strips placed within the standard

load bearing gravel of the street. The rooting volumes ranged from 15 to 30 m³ for Alnus and 45 to 112 113 50 m³ for the *Tilia* trees and spacing between the *Tilia* trees was 15 m, 4-5 m between *Alnus* trees. 114 The underlying soil was clay and subsurface drains were installed on both sites. The level of the water table at the Tilia site was continuously high and the street sloped towards sea at only 2 to 6 m 115 116 above sea level. This site collected water from a larger catchment area compared to the Alnus site, 117 which was fed only with local rainfall. The research sites and soils are described in detail in 118 Riikonen et al. (2011) and Table 1. Trees at both sites were irrigated weekly for the first two years 119 and they remained largely unpruned until late 2008. Thereafter, *Tilia* was pruned annually to achieve necessary crown lifting. All trees had a tree grate that covered a 2.25 m² area installed 120 121 around the tree with open soil underneath. The street surfacing around the Alnus tree grate was 16 x 122 16 cm granite pavers with 10-20 mm drainage gravel grouted joints. The *Tilia* trees, had 10 x 10 cm granite pavers with 10 mm sand grouted joint covered an area of 4 m^2 around the tree grate, and 123 124 outside that, concrete blocks (28 x 16 cm) were installed with 2 mm sand grouted joint.

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126 An automatic monitoring system was built on each soil mixture x tree -combination on each site 127 with a data logger (EnvicLtd., Turku, Finland) to collect data from automated measurements as 128 summarized below. The data used in this study were collected over the 2008-2011 period.

129

130 Soil

Soil volumetric water content (SWC) was measured using ML2x sensors at the depths of 10 and 30
cm from each tree soil surface (DeltaT Devices Ltd., Cambridge, UK, see Riikonen et al. 2011 for
details). The measurement interval was 2 min.

135 Meteorology

136 Air temperature and photosynthetically active radiation (PAR) at the study sites were measured at the height of 8 m (radiation shielded thermistors and QS2 sensors, Delta-T Devices Ltd., 137 138 Cambridge, UK, respectively). In addition to these on-site meteorological measurements, 139 observations from the nearby (4 km) SMEAR III urban measurement station (60°12', 24°57', Järvi 140 et al. 2009) were used. Relative humidity (HMP243, Vaisala Ltd, Vantaa, Finland), precipitation (weighing rain gauge, Ott Pluvio, Kempten, Germany, at 4 min intervals) and wind speed (cup 141 142 anemometer) were measured on the roof of a building (height 52 m), and PAR, global radiation (I) 143 and downward longwave (L^{-}) radiation (PAR lite CNR1, Kipp & Zonen, Delft, Netherlands) from 144 top of a 31 m high measurement tower. Water vapor pressure deficits (D) at the Tilia and Alnus 145 sites were calculated using the local temperatures and relative humidity measured at the SMEAR III 146 station.

147

148 Tree growth, leaf area and tree dimensions

149 Tree trunk circumference at 130 cm height (in April), and tree height, live canopy height and width 150 in two approximately perpendicular directions (in August) were measured with measuring tape each 151 year. Total leaf area of the trees was estimated by fitting a linear regression to the log-transformed 152 branch leaf area and branch basal area (measured in August) and summing up over the over the first 153 degree branches of the tree (Riikonen et al. 2011) (Table 1). Leaf area index (LAI) equalled the 154 total leaf area divided by the projected canopy area. Live crown surface areas were approximated 155 from the canopy dimensions assuming the typical *Tilia* crown shape to be a cone and *Alnus* crown 156 to be an ellipsoid.

157

158 The interannual variation in LAI was derived from shoot measurements. The shoot length increment 159 was measured for 3 shoots each of 6 trees of both species from bud burst to growth cessation with

digital callipers two or three times per week. After the annual shoot and leaf growth had ceased the
leaves on five representative shoots in each of 15 *Tilia* and 22 *Alnus* trees were counted and the
obtained means were used to estimate the remaining number of leaves for the respective species.
The relative number of leaves remaining was that divided by the number of leaves at shoot growth
cessation.

165

166 *Tree transpiration*

167 A self-made (Hölttä et al. 2015) Granier type heat dissipation sensor pair (Granier 1985) was 168 installed in three *Tilia* and three *Alnus* trees nearest to each data logger. Prior to each growing 169 season, holes (3 mm diameter) were drilled at 10 cm apart within the range 50-100 cm height on 170 north and/or east side of the tree trunks. Brass tubes that had been filled with thermally conducting 171 silicone compound were instantly inserted into the holes and sensors (45 mm in length) into the 172 brass tubes. The sensors, measuring the first 45 mm of sapwood inwards from bark, were protected 173 from direct radiation by radiation shields that had been installed inside the trunk guards, allowing 174 free mixing of air. All sensor data obtained from each point were collected by the data loggers at 1-175 10 min intervals.

176

Sap flow at 30-min intervals was calculated according to Granier's (1987) empirical calibration. A five day running mean was used to estimate the temperature difference between the heated and reference sensor at 0-flow conditions (at 2:30-4:30 Eastern European Time, when the lowest diurnal flow rates were recorded) (T_{max}). Measured sap flow is considered to represent tree water use by assuming that the whole of the sapwood cross-section conducted water, which is considered to be a good estimation for young diffuse-porous trees (Gebauer et al. 2008).

184 Most of the sap flow sensors provided continuous data from July to August in 2008, from May to 185 September in 2009 and from May to October in 2010 and 2011. Data obtained before bud break and after leaf fall were omitted. When a tree had more than 1/4 of the 30 min values missing for a day, 186 that day and tree data were excluded from the data set. For a day to be included in the calculation of 187 188 average daily water use, measurements of at least two of the three measured trees per species were 189 required. Days with missing daily sums were gapfilled linearly. The missing daily sums within the 190 used data periods for *Tilia* were in 2008: 9.5%, 2009: 18.9%, 2010: 10.2%, 2011 24.8%, and for 191 Alnus in 2008: 21.8%, 2009: 19.7%, 2010:16.6% and in 2011 10.5%.

192

193 The Penman-Monteith (PM) model

The predicted evapotranspiration (ET_T) for both Tilia and Alnus sites was calculated using a modified PM model (Eq. [1]) in 30 min intervals. The model was optimized against the measured sap flow per projected canopy area (PCA), sf_{meas} (g m⁻² s⁻¹) according to

197
$$c \frac{\Delta R_n + \rho_a C_p(\mathbf{D}) g_a}{(\Delta + \gamma (\mathbf{1} + \frac{g_a}{g_c})) \varphi_v} = ET_T,$$
(Equation 1.)

198 where *c* is the fitted multiplier, Δ is the rate of change of saturated air humidity with temperature 199 (Pa K⁻¹), R_n is the net all-wave radiation (W m⁻²), ρ_a is the dry air density (kg m⁻³), C_p is the 200 specific heat capacity of air (J kg ⁻¹ K⁻¹), *D* is the water vapor pressure deficit (Pa), γ is the 201 psychrometric constant (Pa K⁻¹) and φ_v is the latent heat of vaporization (J g⁻¹). g_a and g_c are the 202 atmospheric and canopy conductances (m s⁻¹). In Eq. [1], g_a is calculated from its inverse, 203 aerodynamic resistance ($r_a = g_a^{-1}$, s m⁻¹) according to

204
$$r_a = \frac{\ln\left(\frac{(z_m - z_d)}{z_{0m}}\right) \ln\left(\frac{(z_m - z_d)}{z_{0v}}\right)}{k^2 u},$$
 (Equation 2.)

where z_m is the height of the measured wind speed (m), z_d is the displacement height (m), z_{0m} and 206 z_{0v} are the roughness lengths for momentum and water vapor (m), k is the von Karman constant and 207 *u* is the wind speed (m s⁻¹). The effect of the atmospheric stability was assumed to be negligible and 208 thus the stability functions for momentum and water vapor were omitted from the calculation of r_a . 209 z_{0m} and z_d were calculated from the mean tree height z_h (0.1 z_h and 2/3 z_h , respectively) and z_{0v} as a 210 rule of thumb $z_{0v} = z_{0m}/10$ (Grimmond and Oke 1991).

211

Models of variable complexity were used to calculate g_c in Eq. [1]. In the simplest model (M1) g_c was a standard parameter (1/70 m s⁻¹) according to FAO guidelines (Allen et al. 1998). In the more complex model (M2) g_c was calculated using a Lohammar-type function that considered the effect of I (W m⁻²) and D (Pa) (Granier et al. 2000):

216
$$g_c = \left(gs_{max} \frac{I}{a+I} \frac{1}{1+bD}\right) L,$$
 (Equation 3.)

where *L* accounts for the effects of leaf area and gs_{max} (m s⁻¹), *a* (W m⁻²) and *b* (Pa⁻¹) are the fitted parameters for each species. The first parameter gives the maximum canopy conductance, *a* is the non-linearity of light modifier (Michaelis-Menten type saturation) and *b* is the vapor pressure deficit modifier (both modifiers obtain values that range from 0 to 1).

221

222 While in M2 L = 1, in the next model M3, L accounts for the effect of leaf area (LA):

223
$$L = \min(1, \frac{1}{6} \text{LAI} \times L_{2}),$$
 (Equation 4.)

where L_s accounts for the change within a growing season. According to Granier et al. (2000), g_c scales linearly with LAI < 6 m² m⁻² and with higher values it saturates in relation to LA. L_s was estimated from mean shoot length growth measurements, assuming the relative leaf area increase was linearly proportional to the relative shoot length increment. After the annual shoot growth had ceased, leaf area scaling followed the relative number of remaining leaves. Additionally in M3, the effect of crown surface area in relation to crown projection area were accounted for by replacing *c* (Eq. [1]) with a term CS/PCA × β . CS refers to crown surface area (m²) and β replaces *c* as a fitting parameter; a separate symbol is used to underline the exclusion of crown dimensions from theparameter.

233

I and R_n used in Eq. [1] and Eq. [3] were indirectly derived. PAR measured at the tree sites (µmol m⁻²) was converted to *I* by using a conversion factor (0.55) that had been derived from 30 min means at the SMEAR III station. Furthermore, the R_n at tree surface was estimated based on Loridan et al. (2010):

238
$$R_n = I(1 - \alpha_0) + \varepsilon_0(L \downarrow -\sigma T^4) - 0.08 I(1 - \alpha_0),$$
 (Equation 5.)

where α_0 and ε_0 are the effective surface albedo and emissivity for deciduous trees, respectively, σ is the Stefan and Boltzmann constant (W m⁻² K⁻⁴) and *T* is the air temperature (K). *R_n* at the street tree sites was further transferred from locally measured PAR by using a conversion factor from the correlation between *R_n* calculated from Eq. [5] and measured PAR at SMEAR III. The *R_n* value that was calculated for the Tilia site was used also for Alnus site, because the PAR sensors at the Alnus site were located in the street canyon and shaded part of the time by the surrounding buildings on its eastern side.

246

247 **PM model fittings**

248 The performance of modified PM model (Eq. [1]) was assessed both annually (annual models) and 249 over the whole 4-year datasets (general models) using all 30 min sap flow values from 2008-2012. 250 The similarity of the acquired PM parameter values in Eq. [1] were compared with Tukey's two-251 sided T test. First, however, a time lag between evapotranspiration, tree sap flux and environmental 252 conditions was accounted for (Granier 1985, Burgess and Dawson 2008). The time lag that gave 253 the best fit (least squares) between the ET_T and sf_{meas} using the whole growing season for each site from 2008 to 2011 (Tables 2 and 3) was chosen and used in all data analyses in this study. The 254 255 tested lag times ranged from 0 to 4 hours and the best fit was obtained for a time lag of 60 min in all years except at the Tilia site in 2009, when the time lag was 30 min, and at the Alnus site in 2010,when time lag was 90 min (Tables 2 and 3).

258

259 Statistical tests for the effect of environmental variables

Comparisons between ET_{T} , *sf_{meas}* and residuals against environmental variables were performed on a daily basis for the annual M3 model with multiple linear regression (SAS procedure REG). A comparison of the daily residuals (*sf_{meas}* - PM ET_T) was made between the SWC and all the meteorological variables in the PM model in backward selection using Akaike's Information Criterion. Both SWC measurement depths were initially included in the model and then the less significant depth was omitted and the fitting was then completed.

266

Furthermore, a linear regression (with logarithmic transformation when necessary) was used to study relationship between daily *D*, *I*, SWC and sap flow density (j_s , g cm⁻² day⁻¹) in July-August each year on the daily sum level, loosely following the approach used by McCarthy and Pataki (2010). To examine the effects of *D* and *I* on j_s , a model

271
$$j_s = i + e \ln(D) + f$$

Ι

(Equation 6.)

with two parameters (e and f) and an intercept (i) was fitted. In the following step the residual (js_{meas} - js_{pred}) was regressed against SWC. The linear regression model was applied with SAS procedure REG and NLIN. Data was fitted separately for each year and site.

275

Measured weekly sap flow sums (calculated per m² of LA) for each year were compared between sites and analyzed by using the Kruskall-Wallis test. All statistical analyses in this study used pvalue of 0.05 as the cut-off point below which differences were considered to be statistically significant.

281 The annual sum of measured *Tilia* tree transpiration in relation to PCA was compared to total 282 annual precipitation measured at the SMEAR III station. Estimates of tree transpiration were made 283 for different canopy cover percentages (proportion of land area covered by horizontal projections of 284 tree crowns) by scaling the measured sap flow per PCA, Similar comparison was also performed 285 using the ET_T estimates obtained from the optimized PM model. The year 2008 was excluded from 286 this analysis due to the high percentage of missing sf_{meas} (65% of daily values over the entire 287 summer for *Tilia* and 57% for *Alnus*). In the estimates accounting for canopy interception, we 288 assumed that 38.2% of precipitation during the growing season and 27% in the wintertime was intercepted by canopy area corresponding to the canopy cover percentage (Xiao and McPherson 289 290 2011).

291 **Results**

292 General meteorological and hydrological conditions

293 The study years 2008-2012 differed from each other especially in regard to their mean daily T and 294 D (Figure 2). July 2010 was clearly warmer (22 °C) than the 30-year mean (18 °C) for Helsinki 295 (Pirinen et al. 2012). Similarly, systematically higher D were measured in July 2010 than in the two 296 previous years. If considered over the entire summer, D was on average higher still in 2011, but this 297 was due to high values in early summer (June). The Tilia site had a higher groundwater table 298 (Riikonen et al. 2011) and higher SWC values than at the Alnus site throughout the entire four-year 299 period. SWC values rose from year to year at the Tilia site, and at the Alnus site, SWC was at its 300 highest also in the last year, 2011 (Figure 2).

301

302 Tree transpiration

- 303 Based on sap flow measurements, *Tilia* water use on a PCA basis in July-August was
- approximately 1/4 of the *Alnus* water use in all studied years (Table 4), but weekly water use per m²

305 of leaf area only showed a significant difference between species in 2009. The mean water use per 306 m^2 of leaf area was at its highest in *Tilia* in 2010 and in *Alnus* in 2009. Approximate daily water use 307 per tree was at its highest in 2010, ranging from 40 to 50 l d⁻¹ for *Tilia* and 60 to 70 l d⁻¹ for *Alnus*. 308

309 The PM model

310 General parameterization

311 To find the optimal PM model parameters to describe street tree sap flow, comparisons between the 312 different models against observations were made. The PM model, fitted to all the study years 313 (general parameterization), explained a minimum of 54% of the 30 min mean transpiration values at 314 its simplest (M1, Table 2). M1 however overestimated transpiration in three years out of the four 315 years studied (bias in the all years model, Table 2). In comparison to M1, the bias in the annual 316 transpiration estimates was reduced considerably (from 6 - 80% to 7 - 56%) in M2. This is 317 attributable to the *I* and *D* modifiers to describe canopy conductance (Eq. [3]) included in M2 318 model. M2 also explained slightly more (61%) of the variation in the transpiration values than M1; 319 the improvement was more notable at the Alnus site. The addition of varying LA and crown surface area to the model (M3) improved the fit further (\mathbb{R}^2 increased by about 0.05, Tables 2 and 3). The 320 M3 model had again smaller bias than M2, but the improvement was not as clear as the one gained 321 322 by the addition of canopy conductance submodel in M2.

323

The parameter *a* describing stomatal response to solar radiation in M3 was thrice as high and gs_{max} less than half for *Tilia* in comparison to *Alnus* (Table 3). These parameter values mean that *Tilia* obtained half of its maximum transpiration rate at more than double the light intensity and had only one third of the maximum transpiration rate in comparison to *Alnus*. The *D* response parameter *b* was only significant for *Alnus*, indicating that for Tilia, additional model to describe stomatal reaction to *D* was not needed. Values of the scaling parameter β in M3 differed little between the

330 species. All the fitted parameters were significantly different for *Tilia* in comparison to *Alnus*331 however.

332

333 Annual parameterization

334 Fitting the models separately for each year (annual parameterization, Tables 2 and 3) improved the 335 explained variation and reduced or removed the trend in the prediction bias in all model versions. In M1 - M2 this was mostly because annual variation was allowed in parameter c, which mainly 336 337 describes the variation in canopy properties. In M1, the R² of individual years varied between 0.48 338 -0.83 with generally better correlation at Tilia than Alnus site. All the models tended to slightly 339 overestimate low sap flow and underestimate high sap flow on both 30 min (Figure 3) and daily 340 (Figure 4) level. Considering the light and D modifiers for canopy conductance (M2) and the 341 within-year leaf area dynamics at annual level (M3, also Figure 3) improved the predictions further 342 $(\mathbb{R}^2 \text{ between } 0.60 - 0.81 \text{ and } 0.67 - 0.86, \text{ respectively})$. On the daily level, the differences between 343 the tested model versions were more pronounced than on the 30 min level (Figure 4).

344

345 The parameters of the g_c function in the annually fitted M3 model were fairly similar within each 346 site in 2009-2011 but with some exceptions they were either at their lowest or highest in 2008 347 (Table 3). Parameter values had less inter-annual variation in *Alnus* than in *Tilia*. The *D* response 348 parameter b was not significant at the Tilia site in 2008 and 2010. The light response parameter a 349 value was at its highest on both sites in 2008 and lowest in 2011. In the models M1-M2 the scaling 350 parameter c increased annually and was always 3-4 times higher at the Alnus site than at the Tilia 351 site. In M3, where crown surface area in relation to projected crown area is taken into account, the 352 scaling parameter β varied less, yet significantly, between sites and years but there was no similar 353 trends as in c and also the differences between the species were reduced.

355 *Residual correlations*

The residual $(sf_{meas} - ET_T)$ correlations with environmental variables (I, D and SWC), used to assess 356 357 the model fit and effect of environmental variables on tree transpiration, ranged from fairly strong $(R^2 = 0.67)$ to non-existent (Table 5). Correlations with environmental variables at the Tilia site 358 359 were weaker than at the Alnus site and at their most obvious in 2008 when the early and late 360 summer measurements were missing. The model predicted higher transpiration than was actually 361 observed in high air temperature at the Alnus site, especially in 2010. Soil water content had a 362 notable role at the Alnus site in all years examined, especially in 2008 and 2011. Soil water content regressed alone against (sf_{meas} - ET_T) was responsible for most of the explanatory power of the 363 364 residual regression model at the Alnus site each year.

365

366 The regression model

367 The regression model (Eq. 6), applied on daily level to find support for PM based analysis of the 368 effect of environmental variables, was significant for both sites and all years except for the Alnus site in 2011. R² varied from 0.48 to 0.88 (Table 6). The model accounted for the observations from 369 370 the Tilia site better than those from the Alnus site. The results showed that D was the more significant driver at the Tilia site in 2008 and Alnus site in 2009, while I dominated at the Tilia site 371 in 2010 and at the Alnus site in 2008. Correlation between SWC and the residuals of the linear 372 373 regression model was significant at the Tilia site only in 2008, but at the Alnus site, it was 374 significant in all years except 2010 (Figure 5).

375

376 Comparison between tree transpiration and rainfall

The measured and modelled transpiration values respective to the annual rainfall in 2009-2011 were estimated with different surface fraction of vegetation scenarios (Table 7) to assess the potential benefits of tree transpiration in managing e.g. stormwater from building roofs. With 20% of the

380 surface covered with *Tilia* crowns, the transpiration from the canopy cover would be 15-20% of the 381 annual rainfall. If the coverage of the canopy would be 70% the respective fraction would be 60-382 80%. For Alnus, the corresponding percentages were much higher, and the annual total rainfall 383 would have been transpired with less than 50% canopy cover in an average year. M1 produced 384 higher transpiration for *Tilia* than what was measured, but the difference was only approximately 385 10%. Correcting for soil water effects (i.e. assuming no water limitation in soil) did not affect the 386 estimates for *Tilia*, but for *Alnus*, the SWC corrected estimates over all three years were slightly 387 higher than the uncorrected ones (not shown).

388

With the columnar *Alnus* trees, transpiration equalling rainfall was reached on average with approximately 35% canopy cover. If we assume that canopy interception accounts for 35-40 % of rainfall during the growing season and 27% in the wintertime (Ilvesniemi et al. 2010, Xiao and McPherson 2011), the estimates of canopy cover that would produce water removal equal to rainfall were about 30% for *Alnus* and 90% for *Tilia*.

394

395 Discussion

396 Tree transpiration

397 In comparison to forest tree transpiration in the boreal zone, street tree transpiration measured at the 398 study sites appeared to be similar or somewhat higher when calculated per LA (Amiro et al. 2006, 399 Arain et al. 2003, Bernier et al. 2006, Daley et al. 2007, Herbst et al. 1999, Launiainen 2010, Oishi 400 et al. 2008). As expected from the reported high values of maximal stomatal conductance of Alnus 401 glutinosa (Eschenbach and Kappen 1999) the water use of Alnus were generally higher than the 402 corresponding values of *Tilia*, although no direct published reference to the behavior of these 403 species in urban conditions have been found. Peters et al. (2010) observed similar or lower 404 transpiration rates for urban *Tilia americana* in Minneapolis compared to *Tilia × vulgaris* in this

study. However, *Tilia americana* is considered to be more shade tolerant than the *Tilia* of this study
(Burns and Honkala 1990) and the late successional species also have a generally lower stomatal
conductivity (Bazzas 1979).

408

409 The heat dissipation method used for sap flow measurements has several known sources of 410 uncertainty (Clearwater et al. 1999). There is a delay in the response of measured sap flow to 411 changes in environmental factors that is caused by both tree- and method-related factors (Tatarinov 412 et al. 2005, Wullschleger et al. 2011). When the relationships between sap flow and transpiration 413 driven by meteorological factors are studied, this delay needs to be taken into consideration. The 414 best fit between tree water use and ET_T in the present study was obtained with a time lag of between 415 30 to 90 min, which is similar to what Granier et al. (2000) reported for coniferous and deciduous 416 forests growing in latitudes that ranged from Canada to the tropics.

417

418 Traditionally it is assumed that no sap flow occurs in predawn hours (see Granier 1987), allowing 419 for T_{max} determination from nocturnal sap flow measurements. It is well known however that this 420 assumption does not always hold true (e.g. Daley and Phillips 2006, Granier 1987, Phillips et al. 421 2003). Despite this, alternative methods for determining T_{max} (with the help of e.g. weather data, 422 Regalado and Ritter 2007) have not yet gained much popularity, and appropriate selection averaging period for T_{max} is the most common method to reduce potential error. If T_{max} in this study 423 424 did not represent true no-flow situation, this leads to underestimation of tree sap flow, which in 425 relation to the role of trees in stormwater management would lead to less critical dimensioning error 426 than the opposite. Also, significant night time evapotranspiration seems to be rare in the boreal 427 environment (Launiainen et al. 2005), reducing further the potential error of incorrect T_{max} 428 determination.

430 The recorded sap flows may not represent tree transpiration accurately because these were 431 measured at the superficial 96-79% of the stem cross-section. We assumed that sap flow was 432 uniform within the whole tree trunk. Although this is not strictly true, the deviations from this with 433 relatively small diffuse porous trees can be expected to be small (Gebauer et al. 2008). The possible 434 overestimation in the measured sap flow based on trunk and sensor dimensions would be 21% at the highest if there were no flow beyond the instrumented sapwood. We could not destructively sample 435 436 our trees and therefore we had no way of knowing whether heartwood was indeed present in our 437 study trees but in relation to tree age and size it was probably not an issue. Tilia sp. have been found 438 to have sapwood lying at lengths beyond the sensor dimensions we used in the present study 439 (Gebauer et al. 2008, Peters et al. 2010, Hölscher 2005). The same reasoning applies to Alnus sp. 440 for which sapwood depth has also been recorded to exceed our sensor length (Moore et al. 2004). 441

442 **The PM model**

443 The adaption of a Lohammar type response of canopy conductance to light and vapor pressure in 444 M2, and its linear-saturating response to leaf area (Granier et al. 2000) in M3 improved the fit 445 between sf_{meas} and ET_T for both general and annual parameterization. General parameterizations of the different PM model versions gave biased predictions over time, but the bias was smaller with 446 447 each consecutive addition of response variables. In M3 the bias in the estimation of annual 448 transpiration was still significant (Tables 2 and 3); the average daily absolute deviation of the 449 predicted value was 35% of the measured value at the Tilia site and 29% at the Alnus site. This 450 would lead to a risk of similarly large over- or underestimation in transpiration prediction made 451 with this model without annual parameterization. In conclusion, satisfactory prediction accuracy 452 with the same parameter values across all four years could not be attained.

453

429

454 Accounting for leaf area and its development within a growing season improved the prediction 455 mainly in spring and autumn, when leaf area was below its annual maximum. The use of shoot 456 growth as a proxy for leaf area development appeared to cause the predicted sap flow to precede the 457 measured sap flow slightly in the spring (data not shown). This was likely because the leaf 458 development from initiation to full maturity lags behind shoot development, and the leaf level 459 functions also vary within year (Xu and Baldocchi 2003; Wilson et al. 2001). We used the approach 460 of saturating the canopy conductance in respect to leaf area index, suggested by Granier et al. 461 (2000). Iterations with our data gave the best fit with slightly higher LAI (~10) than they suggested, 462 implying that saturating leaf area index could be higher in open-grown trees.

463

464 The parameter c was fitted to scale the PM evapotranspiration to sf_{meas} in model versions M1 and M2. When c was allowed to vary from year to year, the prediction bias observed in general 465 466 parameterization (overestimation in 2008 and 2011, underestimation in 2010-2011) was much 467 reduced. In M1, the parameter c parallels the crop coefficient (k_c) to fair extent, although crop 468 coefficient assumes well-watered crop and large uniform canopy (Doorenbos and Pruitt 1977). The 469 uniform canopy assumption is partially met by considering a single-tree crown and the former by 470 excluding the data where obvious soil water limitation was detected. Despite the deviance from 471 original description of crop coefficient, the variation in c gives us an idea about the expected 472 variation of the crop coefficient in urban setting. Divided by LAI, it can be used in urban tree soil 473 volume requirements models (Degaetano 2000; Lindsey and Bassuk 1991). The crop coefficient 474 determined here, calculated to the adjustment factor used by Lindsey and Bassuk (1991), gives 475 clearly lower value for *Tilia* than the default parameters given by the model authors, and somewhat 476 so for *Alnus* as well. Parameter c estimated for *Alnus* was some three times higher than for *Tilia*, 477 resembling the crop coefficient for irrigated fruit crops (Dragoni et al. 2005; Marsal et al. 2013). 478 More of a concern is the high annual variation in *c* in our dataset however. Excluding the variation

resulting from soil water limitation and the year 2008 with its shorter period of measurements still
leaves twofold variation in the value of *c* (Tables 2 and 3). This implies that using a set crop
coefficient for the case study trees could have led to water use estimation errors of similar
magnitude.

483

The PM model, often in a form closely resembling M1 in this study, has been commonly used for orchard trees (e.g. Edwards and Warwick 1984, Fernandez et al. 2001, Nicolas et al. 2005, Pereira et al. 2006). In these studies comparing sap flow and PM estimates, the correlations attained were good and resembled more our annual parameterization than general parameterization efforts. This may be due to e.g. shorter periods of data used or different scaling in terms of leaf or canopy area, which may have prevented detection of possible intra-annual variation in *c*.

490

491 In model M2, the parameter c is not equal to crop coefficient like in M1, because the addition of the 492 explicit canopy conductance function removes the net effect of stomatal regulation from parameter 493 c. The change helped to decrease the model bias but the numerical value of the parameter did not 494 change much between model versions M1 and M2. This suggests that the between year variation in 495 the evapotranspiration levels was related to structural rather than physiological factors. In the model 496 M3, crown surface area is directly included in the model. This improved the model fit and 497 decreased the bias in the general parameterization. Also the differences in the parameter β values 498 and trends between species practically disappeared. This left similar between year variations in the 499 parameter β value in both species, which we were unable to explain with the simple tree crown 500 characterization used in this material.

501

Including crown surface area directly in the model is reasonable for trees growing in an open site,
receiving direct light also from the sun at low angles. The improvement attained in the general

504 parameterization with this addition implies that the single crown architecture needs to be considered 505 in the transpiration estimates as was suggested by Duursma and Mäkelä (2007). Airborne laser 506 scanning as described by Tanhuanpää et al. (2014) shows promise as a means of data collection for 507 a model that includes crown dimensions.

508

509 Considering the crown size and physiology of trees allows a comparison between the different 510 species and provides an explanation for the observed differences in tree water use. Canopy 511 conductance of the Alnus crowns responded more strongly to low light than it did for the Tilia 512 crowns. The Alnus sp. is known to favor moist sites and can have very high transpiration rates 513 (Eschenbach and Kappen 1999). The higher maximal stomatal conductance (parameter gs_{max}) and 514 the more ready responsiveness of stomata to D (parameter b) in Alnus compared to Tilia suggests 515 that in comparison to *Tilia*, *Alnus* has a heavy water use strategy under favourable conditions. On the other hand the light response differences between the species may originate from their very 516 517 different crown architectures, which create different penumbra influences (Stenberg 1998). The 518 columnar Alnus trees have very dense foliage that reduce the light penetration inside the crown 519 cylinder.

520

521 The estimated parameter values give some insight into the observed between year differences in 522 transpiration. The parameters of the canopy conductance part of the M3 model $(g_{s_{max}}, a \text{ and } b)$ 523 suggest higher leaf area specific transpiration in Alnus in 2010 than in 2009 and 2011. This can be 524 explained by gsmax being higher for 2010 than for 2009, and much lower value of a in 2011 than in 525 the previous years. However, both T and D in 2010 were much higher in July-August 2010 than in 2009 or 2011, which reduced the transpiration rate through the lowering of the canopy conductance 526 527 via the D modifier (see Eq. [3]). The D response of the canopy conductance function was similar to 528 that reported by Granier et al. (2000) for several tree species. The negative correlation in (sfmeas -

ET_T) against temperature in 2010 may reflect *D* response but also direct acclimation to prevailing
temperature. High temperatures can directly lower the photosynthesis rate (Smith and Dukes 2013),
which may in turn lead to lower transpiration (Nikinmaa et al. 2013) without any changes in *D*.

533 The notable difference in the parameter values between year 2008 and other years may arise from 534 the poor availability of data during that year. The differences between this and the other years were smaller when the parameters were estimated from the late summer data also in 2009-2011 (data not 535 536 shown). This suggests that the spring and autumn responses of transpiration to the climatic variables 537 differed from each other. Typically, stomatal conductace peaks just before or at leaf full maturity 538 and declines slowly thereafter, but the trend may be difficult to detect due to increasing self-shading 539 (Wang et al. 2014). Annual canopy conductance maximum may dominate more on the 2008 dataset 540 containing only July and August. Perhaps this type of intra-annual variation in leaf properties and 541 different within growing season weather patterns produce the previously mentioned between year 542 differences in parameter β and should thus be considered if general parameterization over several 543 years is attempted.

544

545 Some of the variables examined with the residual $(sf_{meas} - ET_T)$ analysis (I, D and additionally T546 influences D) were included as drivers in the PM model but other variables (SWC) were not. The 547 significance of the former set of variables indicates driver influences that the model formulation 548 cannot fully capture whereas that of SWC describes environmental influence not currently included 549 in the model. The SWC correlation with tree sap flow was confirmed by both analyses in four of the 550 eight year x site combinations, after the effects of meteorological drivers had been taken into 551 account. Lower sap flow during periods of low soil water content was clear at the Alnus site. Data 552 for sap flow in 2008 were only available for July-August when soil water content was fairly low, 553 and SWC effects may have been partially confounded by the fitted parameters. The lower sap flow

in the driest period of 2010 was attributed mainly to *T* and *D*. The *D*, *I* and *T* were all co-correlated and soil was generally driest when the weather was at its warmest in the study period, thus it may merely be a reflection of a tightening stomatal control due to dry soil.

557

558 The regression model

The examination of *sf_{meas}* in comparison to its main environmental drivers *D* and *I* showed that these explained measured sap flow better at the Tilia site than at the Alnus site. The variation explained by the regression model was comparable to street trees in Los Angeles (McCarthy and Pataki 2010). Site differences and interannual variation in the variation explained by the regression model are both in parallel to the PM annual models. The PM model explained some 20% more of the variation than the regression model however, even if the proportion of residuals related to SWC effects was taken into account.

566

The main conclusion from the regression model results is that the finding of SWC related PM model residuals at the Alnus site is supported. Both models agree that SWC effects on tree sap flow were seen in 2008, 2009 and 2011. The effects seen in the PM model residuals also at the Tilia site in 2008 and Alnus site in 2010 were not detected in the regression analysis, which is likely related to the lower level of the effect. Analysis of the PM model residuals is thus understandably stronger tool in the exploration of SWC effects on tree water use.

573

574 Comparison between tree transpiration and rainfall

575 Assessing tree transpiration for larger areas based on the measurements from *Tilia* indicated that 576 annual rainfall and transpiration could be almost equal on years when annual rainfall is close to the

577 local average (660 - 680 mm, Pirinen et al. 2012), if the canopy cover is very high. Annual variation

578 in transpired proportion of rainfall was considerable however. In addition, scaling up the measured

579 canopy transpiration to higher canopy cover is problematic, because the transpiration per tree 580 increases as canopy cover decreases (Oke 1987). Hagishima et al. (2007) found in a plant spacing 581 experiment that transpiration from scattered vegetation was about 2.7 times higher per PCA than 582 from plants placed close together. Accounting for the effects of increasing spacing on transpiration 583 per PCA was not possible with our data, as the trees were planted in rows with nearly even spacing. 584 It appears likely however that because the columnar *Alnus* transpiration model benefited greatly from including the crown surface area, the vertical dimension contributes much in total transpiration 585 586 and thus tighter spacing would notably lower tree transpiration.

587

588 In this study, street tree water use was similar or slightly higher than that of forest trees in 589 comparable climates. It is estimated (Peurasuo et al. 2014) that of all of the trees managed by the 590 city of Helsinki, less than 1% are street trees, and more than 95% grow in environment managed as 591 a forest, which implies that it would be feasible to use transpiration estimates of forest trees for the 592 urban canopy as a whole. Attaining accuracy in smaller scale requires data on street trees as well 593 however, because street trees dominate in the most built-up areas. This is of particular importance 594 because street trees and their transpiration is a potential factor in the urban storm water 595 management. Emerging research shows that in addition to the better known aspect of canopy 596 interception, tree transpiration can indeed contribute considerably to stormwater management of 597 e.g. parking lots (Scharenbroch et al. 2016).

598

599 Knowledge about the behaviour of different tree species in stormwater management structures such 600 as bioswales is just starting to emerge as well (Scharenbroch et al. 2016); findings here underline 601 the importance of species selection. The indication that only 30% canopy cover of *Alnus* could 602 transpire all the annual rainfall indicates quite a good potential for directing rainfall from larger

surface area to their growing media; allowing for example 50% decrease in tree transpiration due to
tighter spacing would still give similar results at about 50% canopy cover.

605

606 Conclusions

607 Our aim was to quantify water use by street trees and analyse how the environmental factors affect 608 it in boreal urban environment. The water use rates of the tree species studied ranged from 20-401 day⁻¹ for *Tilia* and 30-60 l day⁻¹ for *Alnus* and thus were quite dissimilar. According to our 609 610 expectations the Penman-Monteith model could be used to predict the water use of both species 611 accurately. However, this required both parameterization of the transpiration model annually and 612 also allowing the canopy conductance to vary with light and D, and attain different maximal 613 conductance values (different $g_{s_{max}}$) for each year. Accounting for intra-annual leaf area changes 614 and including tree crown dimensions in the model improved the prediction further. The canopy 615 conductance parameterization was overall similar for each species for three out of the four years 616 examined, but the scaling parameters c and β needed to be determined for each year. The PM model 617 appears suitable for estimating urban tree water use, but overt simplifications in the model and 618 parameterization risk considerable biases in the results. Our multi-year observations and PM-model 619 analysis challenges the use of simple crop coefficient for accurate estimation of street tree 620 transpiration.

621

Our analysis of the factors behind the variation in the evapotranspiration showed that the PM model with an additional canopy conductance and leaf area development submodels, scaled to crown surface area, performed satisfactorily in relation to environmental variables included in the model. Soil water status was not included in the model, but was detected to have a significant effect on tree sap flow in several cases. The results of the regression model also indicated that the effects of soil water status on tree transpiration were significant, supporting the need to showing that its inclusionin the predictive models is necessary.

629

The comparison of tree transpiration and precipitation with different canopy cover percentages showed that *Alnus* water use was on average more than 50% of the precipitation with only 20% canopy cover. *Tilia* trees, in comparison, had such a high impact only with canopy cover of 60-70%. Thus it appears that especially the columnar *Alnus* trees have high potential in stormwater management.

635

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857 Figure captions

858 Figure 1. Aerial view of the Viikki distict in Helsinki (©Kaupunkimittausosasto 2011, Helsinki,

Finland). The Tilia site is marked by a cross and the Alnus site is marked by a square.

Figure 2. Daily mean global radiation (I, W m⁻²) measured at the Tilia site and air temperature (T, C°), water vapor pressure deficit (D, kPa) and soil volumetric water content in the fine soil fraction of structural soils (SWC, % vol., mean of 10 and 30 cm measuring depths) at the tree sites (Tilia site: solid light line, Alnus site: dotted black line) during May-October in 2008-2010.

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Figure 3. A comparison of 30 min measured sap flow and predicted annual M3 ET_T at the a) Tilia and b) Alnus site in 2009. Solid line: 1:1 relationship, dashed line: least-squares regression line (Tilia site: y = 0.85x + 0.004 g m⁻² s⁻¹, root-mean-square error (RMSE) 0.013 g m⁻² s⁻¹, R² = 0.83; Alnus site: y = 0.79x + 0.025 g m⁻² s⁻¹, RMSE 0.057 g m⁻² s⁻¹, R² = 0.79). Solid line: 1:1 relationship. Note the difference in scale of the axes.

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Figure 4. Measured daily sap flow and predicted PM ET_T with different model versions (annual parameterization) for the a) Tilia and b) Alnus site in 2009. For Tilia site M1: $R^2 = 0.58$, root-meansquare error (RMSE) 0.75 l m⁻² d⁻¹; M2: R² 0.59, RMSE 0.74 l m⁻² d⁻¹; M3: R² = 0.68, RMSE 0.68 l m⁻² d⁻¹. For Alnus site M1: R² = 0.27, RMSE 3.96 l m⁻² d⁻¹; M2: R² 0.33, RMSE 3.80 l m⁻² d⁻¹; M3: R² = 0.65, RMSE 2.74 l m⁻² d⁻¹. Solid line: 1:1 relationship. Note the difference in scale of the axes.

Figure 5. Daily residual (sf_{meas} - ($i + e \ln (D) + fI$) (Eq. [3]); i= intercept, e and f: fitted parameters) compared to soil water content at the a. Tilia and b. Alnus site in July-August; at each site, the SWC from the measurement depth giving better correlation is shown (the depth of 10 cm for the Tilia site and the depth of 30 cm for the Alnus site). Adjusted R² at the Tilia site in 2008: 0.12, 2009, 2010

- and 2011: not significant; at the Alnus site in 2008: 0.56, 2009: 0.08, 2010: not significant, and in
- 883 2011: 0.13.

886 Tables

- Table 1. Mean characteristics of the trees measured for sap flow in 2008-2011: DBH = trunk
- diameter at breast height (cm), z_t = the tree height (cm), LA = estimated one-sided tree leaf area
- (m^2) , PCA = the projected canopy area (m^2) and LAI = one-sided leaf area index within the crown.
- SD = standard deviation of the mean.

Site	Tilia				Alnus			
Year	2008	2009	2010	2011	2008	2009	2010	2011
DBH±SD	11.1 ± 2.1	11.1 ± 2.4	12.9 ± 2.4	13.9 ± 2.8	12.4 ± 0.4	13.6 ± 0.5	14.7 ± 0.6	16.1 ± 1.1
$z_t \pm SD$	560 ± 80	570 ± 90	600 ± 120	610 ± 120	970 ± 60	1050 ± 50	1130 ± 70	1200 ± 70
LA±SD	52 ± 11	75 ± 25	70 ± 27	65 ± 19	61 ± 17	75 ± 15	100 ± 20	91 ± 18
PCA	9.1 ± 1.7	9.3 ± 2.6	8.9 ± 3.1	10.6 ± 2.9	3.5 ± 1.1	5.1 ± 1.0	4.1 ± 0.4	6.0 ± 1.4
LAI	5.7	8.1	7.9	6.2	17.4	14.17	24.4	15.3

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906 Table 2.

907 The performance and parameters of PM M1 and M2 models with 30 min resolution data. The 908 differences of the predicted transpiration estimates from the measured sums are shown for each 909 year. *c* is the fitting parameter in Eq. [3]. Standard errors (SE) are given with the mean parameter 910 values.

Site	Year	Adj. R ²	$c \pm SE$	Bias (% of meas.)
Tilia M1	2008	0.82	0.25 ± 0.002	12
	2009	0.76	0.36 ± 0.002	13
	2010	0.68	0.49 ± 0.004	12
	2011	0.80	0.22 ± 0.001	17
	All years	0.59	0.35 ± 0.002	2008:56, 2009: 10, 2010: -21, 2011:82
Tilia M2	2008	0.85	0.21 ± 0.004	-3
	2009	0.80	0.37 ± 0.007	-3
	2010	0.71	0.47 ± 0.008	-3
	2011	0.81	0.28 ± 0.005	-1
	All years	0.61	0.35 ± 0.005	2008: 30, 2009: 7, 2010:-33, 2011:56
Alnus M1	2008	0.68	0.98 ± 0.009	11
	2009	0.63	1.24 ± 0.010	12
	2010	0.72	1.84 ± 0.010	8
	2011	0.47	0.92 ± 0.007	4
	All years	0.54	1.27 ± 0.005	2008:44, 2009: 15, 2010: -25, 2011:45
Alnus M2	2008	0.78	0.85 ± 0.024	-4
	2009	0.68	1.34 ± 0.036	0
	2010	0.83	1.58 ± 0.020	-2
	2011	0.60	1.00 ± 0.017	0
	All years	0.61	1.32 ± 0.018	2008: 32, 2009: 6, 2010: -33, 2011:32

Table 3. The performance and parameters of PM M3 model with 30 min resolution data. The differences of the predicted transpiration estimates from the measured sums are shown for each year. *a*, *b*, and gs_{max} are the parameters in Eq. [3] and β replaces *c* in Eq. [3] for model M3. Standard errors (SE) are given with the mean parameter values. Each parameter was compared within species between years with Tukey's t test; values indicated with the same letter (h,j,k,m) do not differ significantly.

Site	Year	Adj. R ²	$a \pm SE$	$b \pm SE$	$gs_{max} \pm SE$	$\beta \pm SE$	Bias
							(% of meas.)
	2008	0.85	267 ± 72 hj	ns	$0.10 \pm$	$0.06 \pm$	-3
	(9 July - 20 Aug.)				0.025 ^k	0.001^{h}	
	2009	0.83	215 ± 20^{h}	$5.2 \cdot 10^{-4} \pm 7.4 \cdot 10^{-1}$	$0.03 \pm$	$0.12 \pm$	-3
	(20 May - 15			5 h	0.003^{h}	0.002^{j}	
	Sept.)						
lia	2010	0.75	$213\pm25\ h^{j}$	ns	$0.03 \pm$	$0.14 \pm$	-5
Ti	(15 May -15				0.003 ^h	0.002 ^k	
	Sept.)						
	2011"	0.86	160 ± 11^{j}	$8.0\cdot10^{-5}\pm$	$0.02 \pm$	$0.08 \pm$	-1
	(14 May-15 Sept.)			3.0·10 ^{-5 j}	0.001 ^j	0.001 ^m	
	All years	0.65	163 ± 12	ns	$0.02 \pm$	$0.11 \pm$	2008:29, 2009: -6,
					0.001	0.001	2010: -31, 2011:53
	2008	0.78	321 ± 87^{h}	$7.8 \cdot 10^{-4} \pm$	$0.16 \pm$	$0.07 \pm$	-4
	(3 July - 23 Aug.)			2.4·10 ^{-4 h}	0.052 ^h	0.002^{h}	
	2009	0.79	58 ± 5 j	$1.1 \cdot 10^{-3} \pm$	$0.04 \pm$	$0.13 \pm$	1
	(5 May - 15 Sept.)			1.2·10 ^{-4 h}	0.003 ^j	0.002^{j}	
S	2010	0.83	61 ± 6^{k}	$1.0 \cdot 10^{-3} \pm$	$0.09 \pm$	$0.12 \pm$	-2
lmu	(19 May - 15			1.6·10 ^{-4 h}	0.011 ^h	0.002 ^k	
A	Sept.)						
	2011	0.67	38 ± 3 ^j	$4.8 \cdot 10^{-3} \pm$	$0.09 \pm$	$0.10 \pm$	0
	(1 May-15 Sept.)			6.7·10 ^{-4 j}	0.012 ^h	0.002 ^m	
	All years	0.72	51 ± 3	$1.5{\cdot}10{\text{-}3} \pm$	$0.06 \pm$	$0.12 \pm$	2008:42, 2009:2,
				$1.1 \cdot 10^{-4}$	0.004	0.001	2010:-22, 2011:24

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- Table 4. Means of high-summer tree water use (1 m⁻² d⁻¹) measured per square meter of projected
- 932 canopy area and leaf area for *Tilia* and *Alnus* trees. Mean water use per tree is given for the
- 933 indicated period and for all measurements available in each year in parentheses. PCA = projected
- 934 canopy area, LA = one-sided leaf area, SD=standard deviation.

	Tree species	Tilia	Tilia	Tilia	Tilia	Alnus	Alnus	Alnus	Alnus
	Year	2008	2009	2010*	2011	2008	2009	2010	2011
	Time period	9 July -	9 July -	9 July-	9 July -	3 July -	3 July -	3 July -	3 July -
	2 - 2 - 2 + 1	20 Aug	20 Aug	30 July	20 Aug	23 Aug	23 Aug	23 Aug	23 Aug
	$l m^{-2} PCA d^{-1} \pm SD$	2.1 ± 0.6	3.4 ± 0.8	5.5 ± 1.9	2.4 ± 0.6	7.9 ± 3.0	13.3 ± 3.5	18.4 ± 3.2	9.6 ± 2.8
	$1 \text{ m}^{-2} \text{ LA } \text{d}^{-1}$	0.37	0.51	0.78	0.38	0.50	1.09	0.87	0.70
	Mean water use per tree, 1 d ⁻¹	19	32 (28)	49 (41)	22 (20)	29	68 (51)	72 (56)	52 (44)
935	* no data for Augu	ust available	e						
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Table 5. Parameter estimates for daily residual (sf_{meas} -PM ET_T) correlations against environmental variables (multiple linear regression) and total R2 for the residual model. T = air temperature (°C), I= global radiation (W m⁻²), D = water vapor pressure deficit (Pa) and SWC = soil water content (% volume). All environmental variables used were daily mean except for I (daily sum of PAR radiation). All parameters except those listed as not significant (ns) had a p value <0.05.

		Tilia site				Alnus site				
Year	2008	2009	2010	2011	2008	2009	2010	2011		
Number of days	42	109	123	125	52	126	119	134		
Intercept	-1.55	ns	ns	ns	-15.26	3.99	23.66	ns		
Т	ns	ns	ns	0.017	-0.27	ns	-0.83	ns		
Ι	-2.9>10-3	ns	ns	ns	ns	ns	ns	-0.022		
М	1.0>10-3	ns	ns	ns	ns	-1.7>10-3	3.2>10-3	ns		
SWC 10 cm	0.10	ns	ns	ns	1.17	ns	ns	0.27		
SWC 30 cm	ns	0.35	ns	ns	ns	-0.29	-0.98	ns		
Adj. R ²	0.34	0.10	0.00	0.02	0.67	0.09	0.05	0.32		
Adj. R ² , SWC as only explanatory variable	0.00	0.10	0.00	0.00	0.61	0.06	0.05	0.15		

			Tili	a site	Alnus site					
	Time	2008	2009	2010	2011	2008	2009	2010	2011	
	period	(9 July -	(1 July -	(1 July -	(1 July -	(3 July -	(1 July -	(1 July -	(1 July -	
		20 Aug.)	31 Aug.)	31 Aug.)	31 Aug.)	23 Aug.)	31 Aug.)	31 Aug.)	31 Aug.)	
	D	*	*		*		*	*		
	Ι		*	*	*	*		*		
	adj. R ²	0.74	0.75	0.52	0.88	0.48	0.48	0.66	ns	
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980										
981										
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977 Table 6. Significant terms in the linear regression $j_s = i + e \ln(D) + f I$, (Eq. [6]), which compares measured

tree transpiration (j_s) to water vapor pressure deficit (D) and global radiation (I). * = $p \le 0.05$.

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999	Table 7. Estimated and measured effective tree water removal in relation to annual rainfall with varying
1000	canopy cover percentages. The estimates were calculated with the least (PM M1 general parameterization)
1001	and most accurate model (PM M3 annual parameterization), and the latter was also corrected for SWC effect
1002	(ETT M3 \times SWChigh/SWCmeas). Additionally, in the last column for each tree species, estimated canopy
1003	interception (37.5% in summer, 27% in winter) is summed to the estimate of water transpired.

		ł	Tilia	009-2011), % of annual rainfall Alnus				
Canopy cover %	Measured transpi- ration	PM M1	PM M3 + SWC correction	PM M3 + SWC corr.+ canopy interception	Measured transpi- ration	PM M1	PM M3 + SWC correction	PM M3 + SWC corr.+ canopy interception
20	16	18	16	22	58	57	65	72
50	41	46	40	56	144	142	163	179
70	57	64	56	78	202	198	228	250



